

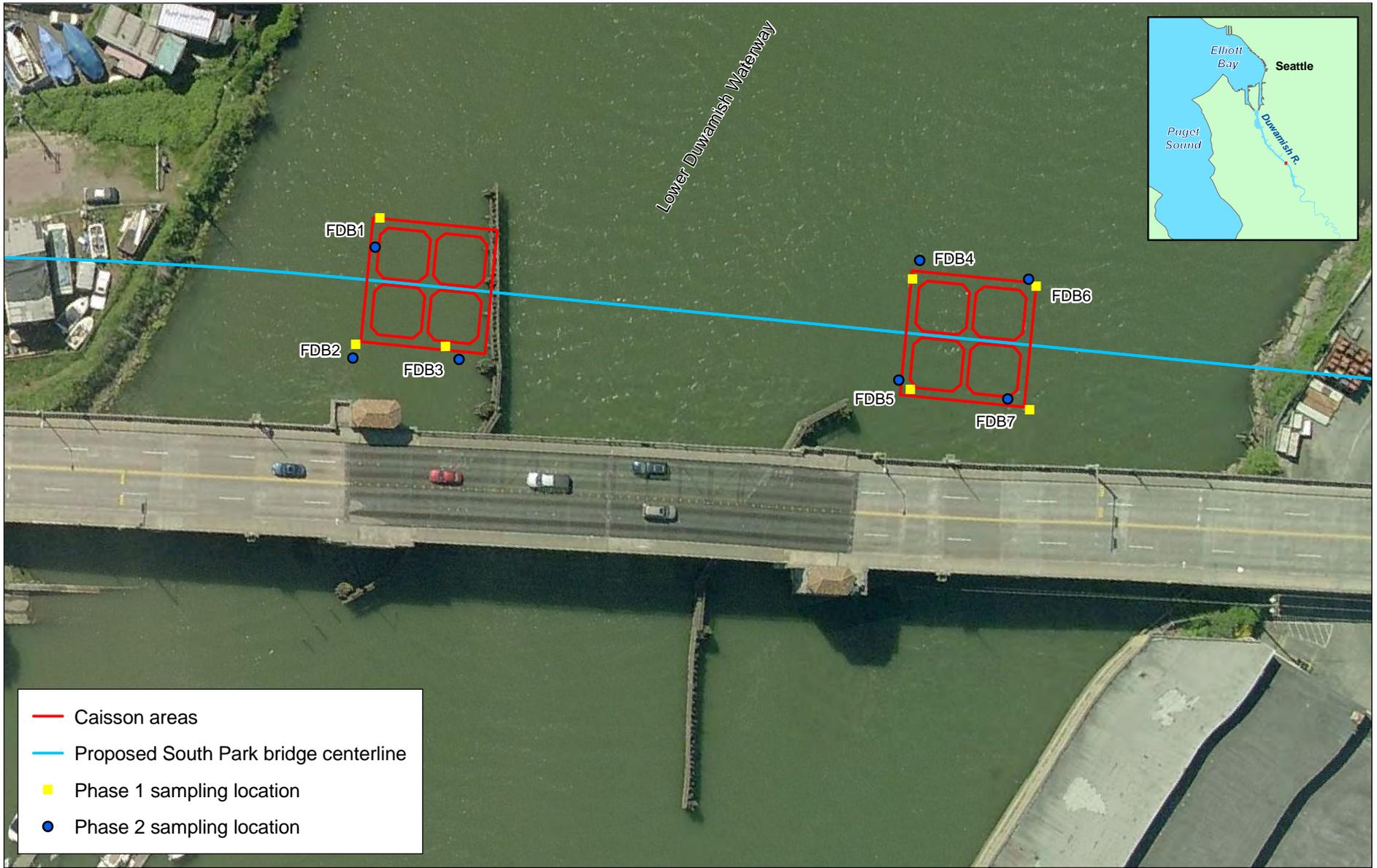
SUBJECT: DETERMINATION ON THE SUITABILITY OF SUPPLEMENTAL CHARACTERIZATION OF PROPOSED DREDGED MATERIAL FROM KING COUNTY DEPARTMENT OF TRANSPORTATION SOUTH PARK BRIDGE PROJECT (CORPS APPLICATION: NWS-2009-1586) IN DUWAMISH RIVER, WASHINGTON EVALUATED UNDER SECTION 404 OF THE CLEAN WATER ACT FOR OPEN-WATER DISPOSAL AT A DMMP NON-DISPERSIVE OPEN-WATER DISPOSAL SITE

1. The following summary reflects the supplemental suitability determination on additional characterization conducted at South Park Bridge Project, and consensus determination of the Dredged Material Management Program (DMMP) agencies (U.S. Army Corps of Engineers, Washington Departments of Ecology and Natural Resources, and the Environmental Protection Agency) on the suitability of an estimated **26,237 cy** of sediment in order to install caissons for the bascule foundation of the replacement bridge evaluated for open-water unconfined disposal at the Elliott Bay non-dispersive open-water disposal site in Seattle, Washington.

Table 1. Project DMMP Tracking Details

| | |
|--|----------------------|
| JARPA APPLICATION NO. | NWS-2009-1586 |
| SAP received | January 13, 2009 |
| SAP approved | |
| Sampling dates: Mud rotary drill rig (Phase 1: Initial characterization) | January 20-27, 2009 |
| Mud rotary drill rig (Phase 2: NB-1 (10-14 ft) for Bioaccumulation Testing) | February 10-12, 2010 |
| Initial Data Characterization Report | June 30, 2010 |
| Revised Characterization Report (Dioxin data corrections, supplemental bioaccumulation data) | August 23, 2010 |
| Recency Determination: High Concern (2 years) | February 2012 |
| DAIS reference number: | SPBRP-1-B-F-295 |

2. **Background.** The South Park Bridge crosses the Lower Duwamish Waterway (LDW) between Tukwila (northern side) and an unincorporated area of King County (southern side). The existing bridge has reached the end of its useful life and must be demolished. King County plans on constructing a new bridge adjacent to the existing bridge after funding is secured. The new bridge requires the installation of deep caissons as part of the bascule foundation for the replacement bridge. The caissons would be approximately 60 ft square and extend to depths of 75 feet for the south bascule and 100 feet for the north bascule (**Figure 1**). Sediment will be dredged in preparation for the caisson installation and during caisson construction, with an estimated total volume of sediment to be dredged from both areas of 26,237 cy.
3. King County Department of Transportation conducted an investigation in 2008 to evaluate the geotechnical characteristics of the sediment in the area of the LDW where the caissons would be installed. Sediment borings were collected and subsequently assessed for sediment quality as a screening level evaluation of potential dredged material disposal alternatives. The results of those analyses were coordinated with the DMMP for review, and feedback from DMMP (**Attachment 1**) were used by WSDOT and their consultants to subsequently design the Sampling and Analysis Plan (SAP) Approach for DMMP characterization articulated in the SAP submitted in January 2009.



- Caisson areas
- Proposed South Park bridge centerline
- Phase 1 sampling location
- Phase 2 sampling location

Figure 1. Sampling locations

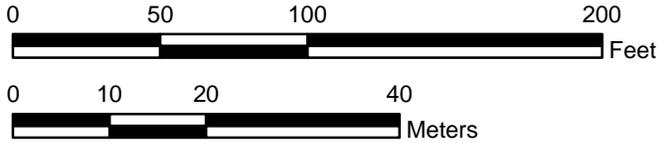


Table 2. Sediment Characterization Details (Errata revision: 12/9/10)

| Location | Depth Interval (ft) Below mudline | DMMU ID | Original Estimated Volume (cy) | Revised Volume Estimate (cy) | Testing Rationale |
|---------------------------------|--|------------|--------------------------------------|---|--|
| South bascule | 0 – 6 | None | 1,100 | 1,150.2 (includes 30.2 cy sand blanket) | Unsuitable based on previous testing results (Wilbur, 2004) |
| | 6 – 10 | SB-1 | 750 | * | DMMP testing (this SDM) |
| | 10 – 14 | SB-2 | 750 | * | DMMP testing (this SDM) |
| | 14 – 18 | SB-3 | 750 | * | Archive pending review of overlying sediment quality |
| | 18 – 75 | None | 8,000 | * | Native (no testing required) |
| | Cofferdam backfill Material to be excavated for caisson construction | None | Omitted | 5,415 (total volume placed: 7,855) | Not tested. Upland material, unsuitable for DMMP disposal |
| | Backfill Material removed prior to cofferdam removal | None | Omitted | 1,040 | Not tested. Upland material, Unsuitable for DMMP disposal |
| North bascule | 0 – 10 | None | 1,850 | 1,897.2 (includes 30.2 cy sand blanket) | Unsuitable based on previous testing results (Wilbur, 2004) |
| | 10 – 14 | NB-1 | 750 | * | DMMP testing (this SDM) |
| | 14 – 18 | NB-2 | 750 | * | Archive pending review overlying sediment quality; PCBs analyzed in Phase II |
| | 18 – 100 | None | 11,500 | * | Native (no testing required) |
| | Cofferdam backfill Material to be excavated for caisson construction | None | Omitted | 5,415 (total volume placed: 7,855) | Not tested. Upland material, unsuitable for DMMP disposal |
| | Backfill Material removed prior to cofferdam removal | None | Omitted | 1,040 | Not tested. Upland material, Unsuitable for DMMP disposal |
| Total estimated volume (Tested) | | | 26,200 | 26,237 | Testing outcome in SDM |
| Backfill Material (Upland) | | | Omitted | 12,910 | Unsuitable for DMMP disposal |

*Estimated volumes are unchanged or lower than the original volume estimates

Table 3. Sampling Coordinates for Phase I and II and Compositing and DMMU Description.

| Location | Core ID | Date | Coordinates | | DMMU ID | Depth (ft) | Analyses |
|-----------------------------|---------|---------|-------------|------------|---------|------------|--|
| | | | Latitude | Longitude | | | |
| Phase I South Bascule | FDB1 | 1/27/09 | 47.529044 | 122.314686 | SB-2 | 10 - 14 | DMMP analytes & Dioxins |
| | FDB2 | 1/27/09 | 47.528989 | 122.314457 | | | |
| | FDB3 | 1/26/09 | 47.529102 | 122.314427 | SB-3 | 14 - 18 | archive |
| Phase I North Bascule | FDB4 | 1/20/09 | 47.529704 | 122.314417 | NB-1 | 10 - 14 | DMMP analytes & Dioxins, toxicity testing |
| | FDB5 | 1/23/09 | 47.529679 | 122.314212 | | | |
| | FDB6 | 1/21/09 | 47.529857 | 122.314367 | | | |
| | FDB7 | 1/22/09 | 47.529825 | 122.314140 | NB-2 | 14 - 18 | Phase I archive; PCBs analyzed Phase II |
| Phase 2 South Bascule | FDB1 | 2/12/10 | 47.529033 | 122.314633 | SB-1 | 6 – 10 | DMMP analytes |
| | FDB2 | 2/12/10 | 47.528983 | 122.314433 | | | |
| | FDB3 | 2/12/10 | 47.529116 | 122.314399 | | | |
| Phase 2 North Bascule | FDB4 | 2/10/10 | 47.529716 | 122.314450 | NB-1 | 10 - 14 | PCBs, bioaccumulation testing |
| | FDB5 | 2/10/10 | 47.529666 | 122.314233 | | | |
| | FDB6 | 2/10/10 | 47.529849 | 122.314383 | | | |
| | FDB7 | 2/10/10 | 47.529800 | 122.314166 | | | |

4. **Additional Details regarding the Bridge Construction.** A 721 cy sand blanket will be placed on top of the contaminated sediments at the pile driving locations prior to installing the work trestles and cofferdams. As noted in **Table 2**, an estimated 30.2 cy of sand blanket (up to 1 foot in depth) would subsequently be removed inside each of the two cofferdams as part of the excavation process and disposed at an approved upland site along with the underlying contaminated material at both bascules. The purpose of the sand blanket is to minimize re-suspension of underlying contaminated sediments, while piles are being driven and later removed.
5. Following the initial excavation within the two 71-ft. by 71-ft. cofferdams, they will be backfilled to depth of approximately 42 ft. to support the initial caisson construction. The volume required will be 15,710 cy (7,855 cy within each cofferdam), which will be upland sourced material. As noted in **Table 2**, 5,415 cy will be removed within each bascule cofferdam for bascule pier construction. Once the caissons are in place, the backfill material remaining between the caissons and the interior walls of the cofferdams will be removed down to the riverbed prior to removing the cofferdam structures. That volume would represent an additional volume of backfill material of 1,040 cy per pier (2,080 cy total). Upland sourced material is not suitable for unconfined open-water disposal.
6. **Sampling and Analysis Plan Submittal/Approval.** The SAP was submitted to the DMMP for review on January 13, 2009 with an anticipated sampling date of January 19, 2009, which did not give the DMMP agencies enough time to formally complete their review and approval of the SAP before sampling was initiated, although DMMP agencies had previously approved the sampling design.
7. **Phase I Sampling.** Sampling was initiated on January 20, 2009 and completed on January 27, 2009 with a mud rotary drill rig during Phase I at three stations within the South Bascule, and at four stations within the North Bascule as summarized in **Tables 2 and 3 (DMMU's SB-2, NB-1, and NB-2)**, and depicted in **Figure 1**.
8. **Phase II Sampling.** Sampling was initiated on February 10, 2010 and completed on February 12, 2010 at the previously occupied three coring stations (mud rotary drill) in the South Bascule, and the previously occupied four coring stations at the North Bascule (**Table 3**). Samples at the **South bascule** were collected and composited for analysis of the 6 – 10 ft layer (**DMMU-SB-1**) for DMMP analytes, excluding dioxin₁, whereas the cores collected within **DMMU-NB-1** (10 – 14 ft layer) at the **North bascule** for bioaccumulation testing.
9. The Phase 1 testing for DMMU's SB-2 and NB-1 included evaluation of dioxins/furans, as well as the PSDDA/DMMP Chemical of Concern list, including TBT. The sampling and analysis plan was generally followed. The sampling and analysis characterization report was initially submitted on June 30, 2010 to the DMMP agencies for review, and a revised report was submitted on August 13, 2010, which corrected dioxin TEQ summaries incorrectly tabulated in earlier report. This report also contained the Phase 2 bioaccumulation testing results as well as NB-1 sediment chemistry for the 14 to 18 ft depths. After reviewing, the DMMP agencies concluded that the quality assurance/quality control guidelines specified by the DMMP were generally complied with, and these data were deemed suitable for decision-making using best-professional-judgment.
10. **Chemical Analysis and Comparison with DMMP Marine Guidelines.** The Agencies' approved sampling and analysis plan was followed and quality assurance/quality control guidelines specified by PSEP and DMMP were generally complied with. A summary of Phase 1 and 2 chemical analysis results for all COC except dioxins/furans is provided in **Table 4**, and demonstrates that all chemicals other than dioxin were either detected or undetected under the DMMP Guidelines except PCBs and Total DDT within **DMMU-NB-1** (10 to 14 ft depth). PCBs in this sample (91 mg/kg-organic carbon normalized) exceeded both SL and Bioaccumulation Trigger in this sample, and DDT was undetected over the SL (28 ug/kg U). Based on these testing results, **DMMU-NB-1**

1 The DMMP decided that dioxin analysis was not required for the Phase 2 testing of DMMU SB-1 (6 – 10 ft below mudline) based on weight-of-evidence in light of low dioxin concentrations observed (0.888 ppb-TEQ) in the Phase 1 SB-2 sample, which represents the underlying sediments (10 - 14ft below mudline) at this location. Additionally, chemical concentrations were low for all other COCs from SB-1.

was first subject to toxicity testing described in **paragraph 14** below, and subsequently to bioaccumulation testing for PCBs described below in **paragraphs 15** through **23**.

11. **Dioxin Testing Results Summary.** Table 5 provides the results of dioxin/furan testing results for two DMMUs as follows: DMMU-SB-2 = 0.685 pptr-TEQ and DMMU-NB-1 = 1.57 pptr-TEQ (U = ½ detection limit).
12. **Dioxin Interim Interpretative Framework.** The DMMP agencies are currently using an interim process for interpreting dioxin data (http://www.nws.usace.army.mil/PublicMenu/Menu.cfm?sitename=DMMO&pagename=Dioxin_Guidelines) pending the development of a programmatic regulatory framework, expected sometime in 2010. The interim guidelines provide a project specific comparison of dioxin/furan concentrations in project dredged material to the disposal site background outside the disposal site. The guidelines applicable to the Elliott Bay non-dispersive disposal site specify the following:
 - a. Comparison of dioxin in test sediments to disposal-site background
 - b. Background is defined using disposal site specific monitoring, which defined an offsite maximum concentration of **12.2 pptr-TEQ**, and an offsite average concentration of **8.7 pptr-TEQ**
 - c. Dioxin concentrations in any given DMMU may not exceed the site maximum (**12.2 pptr-TEQ**)
 - d. Average dioxin concentrations (weighted to the volume of each DMMU) cannot exceed the mean site concentration (**8.7 pptr-TEQ**)
13. **Dioxin Interpretation on Suitability for Unconfined-Open-Water Disposal.** As summarized in paragraph 9 above, DMMU-SB-2 and DMMU-NB-1 were quantitated below the site maximum of **12.2 pptr-TEQ** and below the offsite average of **8.7 pptr-TEQ**, and would be suitable for unconfined open-water disposal at the Elliott Bay site based on these testing results. The volume weighted average for these two DMMUs is **1.13 pptr-TEQ**, which is well below the offsite average of **8.7 pptr-TEQ** (Table 6).
14. **Toxicity Testing Results for DMMU-NB-1.** Table 7a depicts the toxicity testing summary for DMMU-NB-1 due to PCB SL exceedances. Testing was conducted within the 8-week holding time, and the three toxicity tests met all data quality objectives and test acceptability guidelines specified by the DMMP, including control and reference sediment (Carr Inlet) (Table 7b). The bivalve larval test (*Mytilus galloprovincialis*) exhibited a 2-hit response under the non-dispersive site guidelines, and a 1-hit response under the dispersive site guidelines. No other hits were recorded for the other two toxicity tests (e.g., 10-acute toxicity test with *Eohaustorius estuaries*, and the 20-day juvenile polychaete survival and growth bioassay with *Neanthes arenaceodonta*). Therefore, based on these testing results, DMMU-NB-1 is suitable for non-dispersive site disposal at the Elliott Bay site.
15. **Bioaccumulation Testing.** Bioaccumulation testing of DMMU-NB-1 was performed with *Macoma nasuta*, a facultative deposit feeding/suspension feeding bivalve and *Nephtys caecoides*, a burrowing facultative deposit feeding/carnivorous polychaete. The two species were tested together in the same 10-gallon aquaria. The standard PSDDA bioaccumulation test duration is 28 days, but was extended to 45-days to provide a better approximation of steady-state tissue concentrations for the tested chemical (total PCBs) (http://www.nws.usace.army.mil/PublicMenu/documents/DMMO/bioac_00-erratum_09.pdf).
16. As called for in the bioaccumulation protocol, five replicate 10-gallon aquaria were utilized for the negative control, the reference sediment, and for the single tested DMMU-NB-1. Routine water quality metrics (temperature, salinity, dissolved oxygen, pH) were monitored during the exposure period, and the testing conditions employed were Temperature: 15 ± 1°C; Salinity: >25 ‰ ± 2 ‰; Photoperiod: 16 hours/Light: 8 hours/Dark; with gentle aeration to insure that dissolved oxygen does not fall below 40% saturation. All surviving test organisms at the end of the test were depurated 24-hours prior to storing for analysis. During the exposure period, supplemental sediment additions of 0.175 L/week were added to each replicate aquarium, beginning on day 7 of the test. No supplemental feeding of test species was conducted during the 45-day exposure period.

Table 4. South Park Bridge: DMMP Characterization Report Summary

| CHEMICAL NAME | DMMU ID: | | | | | | DMMU-SB-1 (South Bascule) | | | DMMU-SB-2 (South Bascule) | | | DMMU-NB-1 (North Bascule) | | | DMMU-NB-2 (North Bascule) | | | |
|--|----------|--------|--------|--------|----------|-------|----------------------------|---------------|----------|---------------------------|---------------|----------|-----------------------------|---------------|----------|--|---------------|----------|----|
| | DMMP | | | SMS | | | Depth: 6 - 10 ft (Phase 2) | | | Depth: 10-14 ft (Phase 1) | | | Depth: 10 - 14 ft (Phase 1) | | | Depth: 14 - 18 ft (Archived) (Phase 1) | | | |
| | Units | SL | BT | ML | Units | SQS | CSL | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ |
| Antimony | | 150 | | 200 | | | 7.0 | | u | 7.0 | | uj | 6.0 | | uj | NT | | | |
| Arsenic | mg/kg | 57 | 507.1 | 700 | mg/kg | 57 | 93 | 8.0 | | 7.0 | | u | 6.0 | | u | NT | | | |
| Cadmium | mg/kg | 5.1 | 11.3 | 14 | mg/kg | 5.1 | 6.7 | 0.30 | | 0.30 | | u | 0.30 | | u | NT | | | |
| Chromium | mg/kg | (2) | 267 | (2) | mg/kg | 260 | 270 | 18.3 | | 15.9 | | | 12.4 | | | NT | | | |
| Copper | mg/kg | 390 | 1,027 | 1,300 | mg/kg | 390 | 390 | 19.5 | | 18.5 | | | 16.7 | | | NT | | | |
| Lead | mg/kg | 450 | 975 | 1,200 | mg/kg | 450 | 530 | 5.0 | | 7.0 | | | 5.0 | | | NT | | | |
| Mercury | mg/kg | 0.41 | 1.5 | 2.3 | mg/kg | 0.41 | 0.59 | 0.04 | | | | | 0.06 | | u | 0.07 | | | |
| Nickel | mg/kg | 140 | 370 | 370 | mg/kg | -- | -- | 14.0 | | 11.0 | | | 8.0 | | | NT | | | |
| Selenium | mg/kg | (2) | 3 | (2) | mg/kg | -- | -- | 0.70 | | 0.7 | | | 0.7 | | u | NT | | | |
| Silver | mg/kg | 6.1 | 6.1 | 8.4 | mg/kg | 6.1 | 6.1 | 0.4 | | 0.4 | | u | 0.40 | | u | NT | | | |
| Zinc | mg/kg | 410 | 2,783 | 3,800 | mg/kg | 410 | 960 | 39.0 | | 40.0 | | | 30.0 | | | NT | | | |
| TBT ion (porewater) | ug/L | 0.15 | 0.15 | | ug/L | 0.05 | 0.35 | 0.008 | | 0.008 | | u | 0.008 | | u | NT | | | |
| Naphthalene | ug/kg | 2,100 | | 2,400 | mg/kg-OC | 99 | 170 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Acenaphthylene | ug/kg | 2,000 | | 2,000 | mg/kg-OC | 66 | 66 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Acenaphthene | ug/kg | 500 | | 2,000 | mg/kg-OC | 16 | 57 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Fluorene | ug/kg | 540 | | 3,600 | mg/kg-OC | 23 | 79 | 55.0 | 4.8 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Phenanthrene | ug/kg | 1,500 | | 2,100 | mg/kg-OC | 100 | 480 | 19.0 | 1.7 | u | 21.0 | 1.94 | u | 14.0 | 1.4 | j | NT | | |
| Anthracene | ug/kg | 560 | | 13,000 | mg/kg-OC | 220 | 1,200 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| 2-Methylnaphthalene | ug/kg | 670 | | 1,900 | mg/kg-OC | 38 | 64 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Total LPAH | ug/kg | 5,200 | | 29,000 | mg/kg-OC | 370 | 780 | 55.0 | 4.8 | u | 21.0 | 1.94 | u | 14.0 | 1.4 | u | NT | | |
| Fluoranthene | ug/kg | 1,700 | 4,600 | 30,000 | mg/kg-OC | 160 | 1,200 | 19.0 | 1.7 | u | 39.0 | 3.61 | u | 36.0 | 3.6 | u | NT | | |
| Pyrene | ug/kg | 2,600 | 11,980 | 16,000 | mg/kg-OC | 1,000 | 1,400 | 44.0 | 3.9 | u | 34.0 | 3.15 | u | 61.0 | 6.1 | u | NT | | |
| Benzo(a)anthracene | ug/kg | 1,300 | | 5,100 | mg/kg-OC | 110 | 270 | 19.0 | 1.7 | u | 13.0 | 1.20 | j | 21.0 | 2.1 | u | NT | | |
| Chrysene | ug/kg | 1,400 | | 21,000 | mg/kg-OC | 110 | 460 | 19.0 | 1.7 | uj | 15.0 | 1.39 | j | 27.0 | 2.7 | u | NT | | |
| Total Benzo(b+k)fluoranthenes | ug/kg | 3,200 | | 9,900 | mg/kg-OC | 230 | 450 | 19.0 | 1.7 | uj | 19.0 | 1.76 | u | 52.0 | 5.2 | j | NT | | |
| Benzo(a)pyrene | ug/kg | 1,600 | | 3,600 | mg/kg-OC | 99 | 210 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 24.0 | 2.4 | u | NT | | |
| Indeno(1,2,3-cd)pyrene | ug/kg | 600 | | 4,400 | mg/kg-OC | 34 | 88 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 10.0 | 1.0 | j | NT | | |
| Dibenzo(a,h)anthracene | ug/kg | 230 | | 1,900 | mg/kg-OC | 12 | 33 | 24.0 | 2.1 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Benzo(g,h,i)perylene | ug/kg | 670 | | 3,200 | mg/kg-OC | 31 | 78 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 12.0 | 1.2 | j | NT | | |
| Total HPAH | ug/kg | 12,000 | | 69,000 | mg/kg-OC | 960 | 5,300 | 68.0 | 6.0 | u | 62.0 | 5.74 | j | 233 | 23.3 | j | NT | | |
| 1,3-Dichlorobenzene | ug/kg | 170 | | | mg/kg-OC | 2.3 | 2.3 | 1.0 | 0.1 | u | 1.4 | 0.13 | u | 1.2 | 0.12 | u | NT | | |
| 1,4-Dichlorobenzene | ug/kg | 110 | | 120 | mg/kg-OC | 3.1 | 9 | 1.0 | 0.1 | u | 1.4 | 0.13 | u | 1.2 | 0.12 | u | NT | | |
| 1,2-Dichlorobenzene | ug/kg | 35 | | 110 | mg/kg-OC | 2.3 | 2.3 | 1.0 | 0.1 | u | 1.4 | 0.13 | u | 1.2 | 0.12 | u | NT | | |
| 1,2,4-Trichlorobenzene | ug/kg | 31 | | 64 | mg/kg-OC | 0.81 | 1.8 | 5.2 | 0.5 | u | 6.8 | 0.63 | u | 5.8 | 0.58 | u | NT | | |
| Hexachlorobenzene (HCB) | ug/kg | 22 | 168 | 230 | mg/kg-OC | 0.38 | 2.3 | 0.96 | 0.1 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Dimethylphthalate | ug/kg | 71 | | 1,400 | mg/kg-OC | 53 | 53 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Diethylphthalate | ug/kg | 200 | | 1,200 | mg/kg-OC | 61 | 110 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Di-n-butylphthalate | ug/kg | 1,400 | | 5,100 | mg/kg-OC | 220 | 1,700 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Butylbenzylphthalate | ug/kg | 63 | | 970 | mg/kg-OC | 4.9 | 64 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 13.0 | 1.3 | j | NT | | |
| Bis(2-ethylhexyl)phthalate | ug/kg | 1,300 | | 8,300 | mg/kg-OC | 47 | 78 | 19.0 | 1.7 | u | 67.0 | 6.20 | u | 280.0 | 28.0 | u | NT | | |
| Di-n-octylphthalate | ug/kg | 6,200 | | 6,200 | mg/kg-OC | 58 | 4,500 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Phenol | ug/kg | 420 | | 1,200 | ug/kg | 420 | 1,200 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| 2-Methylphenol | ug/kg | 63 | | 77 | ug/kg | 63 | 63 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| 4-Methylphenol | ug/kg | 670 | | 3,600 | ug/kg | 670 | 670 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| 2,4-Dimethylphenol | ug/kg | 29 | | 210 | ug/kg | 29 | 29 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Pentachlorophenol | ug/kg | 400 | | 690 | ug/kg | 360 | 690 | 97.0 | 8.5 | u | 97.0 | 8.98 | u | 99.0 | 9.9 | u | NT | | |
| Benzyl alcohol | ug/kg | 57 | | 87 | ug/kg | 57 | 73 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Benzoic acid | ug/kg | 650 | | 760 | ug/kg | 650 | 650 | 190.0 | 16.7 | u | 190.0 | 17.6 | u | 200.0 | 20.0 | u | NT | | |
| Dibenzofuran | ug/kg | 540 | | 1,700 | mg/kg-OC | 15 | 58 | NT | | NT | | | NT | | | NT | | | |
| Hexachloroethane | ug/kg | 600 | | 1,600 | mg/kg-OC | | | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Hexachlorobutadiene | ug/kg | 29 | | 270 | mg/kg-OC | 3.9 | 6.2 | 0.96 | 0.1 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| N-Nitrosodiphenylamine | ug/kg | 280 | | 130 | mg/kg-OC | 11 | 11 | 19.0 | 1.7 | u | 19.0 | 1.76 | u | 20.0 | 2.0 | u | NT | | |
| Trichloroethene | ug/kg | 160 | | 1,600 | ug/kg | -- | -- | 1.0 | | u | 1.4 | | u | 1.2 | | u | NT | | |
| Tetrachloroethene | ug/kg | 57 | | 210 | ug/kg | -- | -- | 1.0 | | u | 1.4 | | u | 1.2 | | u | NT | | |
| Ethylbenzene | ug/kg | 10 | | 50 | ug/kg | -- | -- | 1.0 | | u | 1.4 | | u | 1.2 | | u | NT | | |
| Total Zylene (sum of o-,m-,p-) | ug/kg | 40 | | 160 | ug/kg | -- | -- | 1.0 | | u | 1.4 | | u | 1.2 | | u | NT | | |
| 4,4'-DDE | ug/kg | | | | | | | 1.9 | | u | 2.0 | | u | 8.8 | | u | NT | | |
| 4,4'-DDD | ug/kg | | | | | | | 1.9 | | u | 2.0 | | u | 8.8 | | u | NT | | |
| 4,4'-DDT | ug/kg | | | | | | | 1.9 | | u | 2.0 | | u | 28.0 | | u | NT | | |
| Total DDT (sum of 4,4'-DDD, 4,4'-DDE and 4,4'-DDT) | ug/kg | 6.9 | 50 | 69 | | -- | -- | 1.9 | | u | 2.0 | | u | 28.0 | | u | NT | | |
| Aldrin | ug/kg | 10 | | | | -- | -- | 1.9 | | u | 0.98 | | u | 4.4 | | u | NT | | |
| Chlordane | ug/kg | 10 | 37 | | | -- | -- | 0.96 | | u | 0.98 | | u | 4.4 | | u | NT | | |
| Dieldrin | ug/kg | 10 | | | | -- | -- | 1.9 | | u | 2.0 | | u | 8.8 | | u | NT | | |
| Heptachlor | ug/kg | 10 | | | | -- | -- | 0.96 | | u | 0.98 | | u | 4.4 | | u | NT | | |
| Alpha-BHC | ug/kg | | 10 | | | -- | -- | 0.96 | | u | 0.98 | | u | 4.4 | | u | NT | | |
| Gamma-BHC (Lindane) | ug/kg | 10 | | | | -- | -- | 0.96 | | u | 0.98 | | u | 4.4 | | u | NT | | |
| Aroclor 1016 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 88.0 | | u | 19.0 | | u |
| Aroclor 1221 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 88.0 | | u | 19.0 | | u |
| Aroclor 1232 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 88.0 | | u | 19.0 | | u |
| Aroclor 1242 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 88.0 | | u | 19.0 | | u |
| Aroclor 1248 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 110.0 | | u | 19.0 | | u |
| Aroclor 1254 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 360.0 | | u | 19.0 | | u |
| Aroclor 1260 | ug/kg | | | | | | | 3.9 | | u | 19.0 | | u | 550.0 | | u | 19.0 | | u |
| Total PCBs | ug/kg | 130 | | | | | | | | | | | | | | | | | |

Table 4. South Park Bridge: DMMP Characterization Report Summary

| CHEMICAL NAME | Units | DMMP | | | Units | DMMU ID: | | | DMMU-SB-1 (South Bascule) | | | DMMU-SB-2 (South Bascule) | | | DMMU-NB-1 (North Bascule) | | | DMMU-NB-2 (North Bascule) | | | | |
|--|-------|------|----|----|-------|----------|-----|----------------------------|---------------------------|-----|---------------------------|---------------------------|-----|-----------------------------|---------------------------|------------|--|---------------------------|----|---------------------------|------------|--|
| | | SL | BT | ML | | SMS | | Depth: 6 - 10 ft (Phase 2) | | | Depth: 10-14 ft (Phase 1) | | | Depth: 10 - 14 ft (Phase 1) | | | Depth: 14 - 18 ft (Archived) (Phase 1) | | | | | |
| | | | | | | SQS | CSL | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ | mg/kg-dry wgt | mg/kg-OC | VQ | | | |
| | | | | | | DMMP | SMS | | DMMP | SMS | | DMMP | SMS | | DMMP | SMS | | | | | | |
| Dioxin (TEQ: see Table 6 for detailed results) | ng/kg | | | | | | | | | | 0.685 | | | | | 1.57 | | | | | | |
| Total Solids | % | | | | | | | | | | 71.0 | | | | | 68.1 | | | | | 75.6 | |
| Total Volatile Solids | % | | | | | | | | | | 3.06 | | | | | 2.7 | | | | | NT | |
| Total Organic Carbon | % | | | | | | | | | | 1.14 | | | | | 1.08 | | | | | 0.79 | |
| Total Ammonia | mg/kg | | | | | | | | | | 44.6 | | | | | 50.4 | | | | | 6.45 | |
| Total Sulfides | mg/kg | | | | | | | | | | 308.0 | | | | | 384.0 | | | | | 63.4 | |
| Gravel | % | | | | | | | | | | 0.8 | | | | | 3.3 | | | | | 1.2 | |
| Sand | % | | | | | | | | | | 75.5 | | | | | 64.6 | | | | | 90.8 | |
| Silt | % | | | | | | | | | | 20.7 | | | | | 25.4 | | | | | 5.3 | |
| Clay | % | | | | | | | | | | 3.6 | | | | | 6.6 | | | | | 2.6 | |
| Fines (percent silt + clay) | % | | | | | | | | | | 24.3 | | | | | 32.0 | | | | | 7.9 | |
| Eohaustorius estuarinus hits: | | | | | | | | | | | | | | | | | | | | | NH | |
| Mytilus galloprovincialis hits: | | | | | | | | | | | | | | | | | | | | | 2H | |
| Neanthes arenaceodentata hits: | | | | | | | | | | | | | | | | | | | | | NH | |
| Bioassay Determination: (P/F) | | | | | | | | | | | NA | | | | | NA | | | | | PASS (ND) | |
| BTs exceeded: | | | | | | | | | | | No | | | | | No | | | | | No | |
| Bioaccumulation conducted: | | | | | | | | | | | No | | | | | No | | | | | No | |
| Bioaccumulation Determination: (P/F) | | | | | | | | | | | No | | | | | No | | | | | PASS (BPJ) | |
| ML Rule exceeded: | | | | | | | | | | | No | | | | | No | | | | | No | |
| PSDDA Determination: | | | | | | | | | | | PASS | | | | | PASS | | | | | PASS (ND) | |
| DMMU Volume: | cy | | | | | | | | | | 750 | | | | | 750 | | | | | 750 | |
| Rank | | | | | | | | | | | | | | | | H | | | | | H | |
| Mean Core sampling depth | ft | | | | | | | | | | 4' | | | | | 4' | | | | | 4' | |
| Maximum sampling depth (mudline) (with Z-sample) | ft | | | | | | | | | | 6' to 10' | | | | | 10' to 14' | | | | | 10' to 14' | |
| DMMU ID: | | | | | | | | | | | DMMU-SB-1 (South Bascule) | | | DMMU-SB-2 (South Bascule) | | | DMMU-NB-1 (North Bascule) | | | DMMU-NB-2 (North Bascule) | | |

Legend:

- SL = Screening Level exceedance
- BT = Bioaccumulation Trigger exceedance
- P = Pass (Suitable for UCOWD)
- SQS = Sediment Quality Standards exceedance (SMS)
- CSL = Cleanup Screening Level exceedance (SMS)
- NH = No Hit
- NT = Not Tested
- 2H = 2 Hit Response
- ND = Nondispersive site Disposal

- VQ = Validation Qualifier
- UCOWD = Unconfined open-water disposal
- u = undetected at the reporting limit
- uj = result undetected at the estimated reporting limit shown
- j = Estimated Concentration (< reporting limit)

Table 5. South Park Bridge Dioxin Testing Results Summary

| Dioxin/furan | WHO (05) TEF | DMMU ID | DMMU-SB-2 (South Bascule) | | | DMMU-NB-1 (North Bascule) | | |
|--------------------------|-----------------|--------------|---------------------------|----|-----------------|---------------------------|----|-------------------|
| | | SAMPLE DEPTH | -10 to 14 ft (MLLW) | | | -10 to 14 ft (MLLW) | | |
| | | UNIT | DMMU-S1 | LQ | TEQ | DMMU-S2 | LQ | TEQ |
| 2,3,7,8-TCDD 1 | | ng/kg dw | 0.0446 | u | 0.0223 | 0.0811 | | 0.0811 |
| 1,2,3,7,8-PeCDD 1 | | ng/kg dw | 0.223 | u | 0.1115 | 0.222 | u | 0.111 |
| 1,2,3,4,7,8-HxCDD 0.1 | | ng/kg dw | 0.223 | u | 0.01115 | 0.222 | u | 0.0111 |
| 1,2,3,6,7,8-HxCDD 0.1 | | ng/kg dw | 0.293 | u | 0.01465 | 1.21 | | 0.121 |
| 1,2,3,7,8,9-HxCDD 0.1 | | ng/kg dw | 0.223 | u | 0.01115 | 0.523 | | 0.0523 |
| 1,2,3,4,6,7,8-HpCDD 0.01 | | ng/kg dw | 6.48 | | 0.0648 | 33.7 | | 0.337 |
| OCDD 0.0003 | | ng/kg dw | 62.8 | | 0.01884 | 275 | | 0.0825 |
| 2,3,7,8-TCDF 0.1 | | ng/kg dw | 0.116 | u | 0.0058 | 0.201 | | 0.0201 |
| 1,2,3,7,8-PeCDF 0.03 | | ng/kg dw | 0.223 | u | 0.003345 | 0.222 | u | 0.00333 |
| 2,3,4,7,8-PeCDF | 0.3 | ng/kg dw | 0.612 | | 0.1836 | 0.787 | | 0.2361 |
| 1,2,3,4,7,8-HxCDF | 0.1 | ng/kg dw | 1.03 | | 0.103 | 2.06 | | 0.206 |
| 1,2,3,6,7,8-HxCDF | 0.1 | ng/kg dw | 0.263 | u | 0.01315 | 0.553 | | 0.0553 |
| 1,2,3,7,8,9-HxCDF | 0.1 | ng/kg dw | 0.476 | | 0.0476 | 0.786 | | 0.0786 |
| 2,3,4,6,7,8-HxCDF | 0.1 | ng/kg dw | 0.223 | u | 0.01115 | 0.288 | u | 0.0144 |
| 1,2,3,4,6,7,8-HpCDF | 0.01 | ng/kg dw | 5.62 | | 0.0562 | 12.8 | | 0.128 |
| 1,2,3,4,7,8,9-HpCDF | 0.01 | ng/kg dw | 0.407 | | 0.00407 | 1.38 | | 0.0138 |
| OCDF | 0.0003 | ng/kg dw | 7.93 | | 0.002379 | 59.8 | | 0.01794 |
| Total TEQ: (u = 1/2) | | | | | 0.685 | | | 1.57 |
| Total TEQ: (u = 0) | | | | | 0.48 | | | 1.43 |
| Total TOC, %: | | | | | 1.08 | | | 1.0, 0.736 |

LQ = Laboratory Qualifier

U = not detected at given concentration

Table 6. South Park Bridge Volume Weighted Average (VWA) Dioxin Concentrations

| DMMU Core ID | Depth, ft | Volume (CY) | TCDD/F TEQ | ng/kg-dw | Product (Vol x TEQ) | ng x cy/kg x DMMU | Product/total | Proportional contribution/Suitable DMMU |
|---------------------------|-------------|-----------------|------------|----------|---------------------|-------------------|---------------|---|
| SB-2 | -10 to -14' | 750 | 0.685 | ng/kg-dw | 514 | ng x cy/kg | 30% | % of Total DMMU |
| NB-1 | -10 to -14' | 750 | 1.57 | ng/kg-dw | 1,178 | ng x cy/kg | 70% | % of Total DMMU |
| Totals (Suitable): | | 1,500 cy | | | 1,691 | ng x cy/kg | 1.13 | ng/kg-dw/Project (VWA) |

Table 7a. Bioassay Testing Results Summary for DMMU-NB-1 (10-14 ft)

| BIOASSAY EVALUATION GUIDELINES | AMPHIPOD TEST | | | LARVAL TEST | | | NEANTHES TEST | | | DMMP Determination |
|---|------------------------|---------|--------|----------------------------------|---------|--------|----------------------------|---------|--------|--|
| | CRITERIA | RESULTS | STATUS | CRITERIA | RESULTS | STATUS | CRITERIA | RESULTS | STATUS | |
| Negative control performance standard | | 0% | Pass | $Nc/I \geq 0.70$ | 0.99 | Pass | $Mc \leq 10\%$ | 0% | Pass | |
| Reference sediment performance standard | $Mr-Mc \leq 20\%$ | 3% | Pass | $Nr/Nc \geq 0.65$ | 0.84 | Pass | $MIGc \geq 0.38$ | 0.75 | Pass | |
| | | | | | | | $Mr \leq 20\%$ | 4% | Pass | |
| | | | | | | | $MIGr/MIGc \geq 0.80$ | 1.1 | Pass | |
| Dispersive disposal | $Mt - Mc > 20\%$ | 1.0% | Pass | $Nt/Nc < 0.80$ | 0.585 | Yes | $MIGt/MIGc < 0.80$ | 1.1 | Pass | Not Suitable Dispersive Site Disposal |
| | Mt vs Mr SS (p = 0.05) | Not SS | Not SS | Nt/Nc vs Nr/Nc SS (p = 0.10) | SS | SS | MIGt vs MIGc SS (p = 0.05) | SS | SS | |
| | $Mt - Mr > 10\%$ | 2.0% | Pass | $Nr/Nc - Nt/Nc > 0.15$ | 0.26 | 1-Hit | $MIGt/MIGr < 0.70$ | 1.0 | Pass | |
| Non-dispersive disposal | $Mt - Mc > 20\%$ | 1.0% | Pass | $Nt/Nc < 0.80$ | 0.585 | Yes | $MIGt/MIGc < 0.80$ | 1.1 | Pass | Suitable Non-Dispersive Site Disposal |
| | Mt vs Mr SS (p = 0.05) | Not SS | Not SS | Nt/Nc vs Nr/Nc SS (p = 0.10) | SS | SS | MIGt vs MIGc SS (p = 0.05) | SS | SS | |
| | $Mt - Mr > 30\%$ | 2.0% | Pass | $Nr/Nc - Nt/Nc > 0.30$ | 0.26 | 2-Hit | $MIGt/MIGr < 0.50$ | 1.0 | Pass | |

Legend.

M = mortality

N = normal larvae

I = Initial Count

MIG = mean individual growth rate

Table 7b. – DMMP EVALUATION GUIDELINES (BIOASSAYS)

| Bioassay | Negative Control Performance Standard | Reference Sediment Performance Standard | Dispersive Disposal Site Interpretation Guidelines | | Nondispersive Disposal Site Interpretation Guidelines | |
|------------------------|--|---|--|------------|--|----------------------|
| | | | 1-hit rule | 2-hit rule | 1-hit rule | 2-hit rule |
| Amphipod | $M_C \leq 10\%$ | $M_R - M_C \leq 20\%$ | $M_T - M_C > 20\%$ and M_T vs M_R SD ($p=.05$) and | | $M_T - M_C > 20\%$ and M_T vs M_R SD ($p=.05$) and | |
| | | | $M_T - M_R > 10\%$ | NOCN | $M_T - M_R > 30\%$ | NOCN |
| Sediment Larval | $N_C \div I \geq 0.70$ | $N_R \div N_C \geq 0.65$ | $N_T \div N_C < 0.80$ and N_T/N_C vs N_R/N_C SD ($p=.10$) and | | $N_T \div N_C < 0.80$ and N_T/N_C vs N_R/N_C SD ($p=.10$) and | |
| | | | $N_R/N_C - N_T/N_C > 0.15$ | NOCN | $N_R/N_C - N_T/N_C > 0.30$ | NOCN |
| <i>Neanthes</i> growth | $M_C \leq 10\%$ $MIG \geq 0.38$ mg/ind/day | $MIG_R \div MIG_C \geq 0.80$ | $MIG_T \div MIG_C < 0.80$ and MIG_T vs MIG_R SD ($p=.05$) and | | $MIG_T \div MIG_C < 0.80$ and MIG_T vs MIG_R SD ($p=.05$) and | |
| | | | $MIG_T/MIG_R < 0.70$ | NOCN | $MIG_T/MIG_R < 0.50$ | $MIG_T/MIG_R < 0.70$ |

M = mortality, N = normals, I = initial count, MIG = mean individual growth rate,
 SD = statistically different, NOCN = no other conditions necessary, N/A = not applicable
 Subscripts: R = reference sediment, C = negative control, T = test sediment

17. The following deviations from the standard DMMP bioaccumulation testing protocol occurred and **were not brought to the attention of the DMMP agencies** until the conclusion of the test:
- a) In resampling to collect the sediment required to conduct the bioaccumulation test for DMMU-NB-1, the applicant failed to collect sufficient volume of test sediment as prescribed by the bioaccumulation protocol. Therefore, the testing laboratory (Northwestern Aquatic Sciences) utilized 3.4 L of sediment/aquaria, rather than the 4.0 L / aquaria called for by the protocol.
 - b) As a result of the lower volume of sediment available for bioaccumulation testing, the initial stocking density of both test species (*Macoma*, *Nephtys*) was lower than recommended by the protocol (*Macoma*: stocked with 12 clams rather than 15 recommended; *Nephtys*: stocked with 20 worms rather than 60 recommended by protocol).
 - c) The lower stocking density combined with weight loss during the 45-day testing exposure period resulting in insufficient biomass for both species to accomplish the 5-replicate analysis per treatment (e.g., control, reference, test sediment) recommended by the bioaccumulation protocol. For *Nephtys*, there was sufficient biomass to conduct only single un-replicated analyses of initial and 45-day control, reference, and test sediment. For *Macoma*, 3 replicate analyses were conducted for Day 0 (control), 45-Day control, and test sediment, with sufficient biomass for the full five replicate analyses for the *Macoma* reference treatment.
18. Survival was generally good as shown in **Attachment 2** for both test species, with higher survival noted for *Nephtys*. **Attachment 3** summarizes the initial and final biomass for both *Macoma* and *Nephtys* on day 0 and day 45 of the test, and both species lost weight during the exposure period, with *Macoma* averaging 79% of the initial weight. Forty-five-day *Nephtys* lost more weight compared to *Macoma*, weighing 51% of the initial worm weight in controls. Weight losses observed in reference and treatment (**NB-1**) were 67% and 62%, respectively, as compared to the initial weights, which were somewhat higher than the control worms.
19. **Tissue Chemistry**. **Attachment 4** provides the full PCB Aroclor analysis results for both species for PCB tissue analyses. As noted earlier, there was insufficient biomass for both species at the end of the 45-day exposure period to conduct the full five replicate analyses per treatment (e.g., control, reference, test sediment) as required by the DMMP bioaccumulation protocol.
20. Tissue concentrations of PCB from the 45-day exposures were compared statistically to the appropriate reference sediment for *Macoma* only. As noted in **Table 8**, the calculated ratios of initial to retested sediment PCB concentrations were used to adjust the observed tissue concentrations, as the retested PCBs were lower than the initial result by a 5.1 ratio. Statistical comparisons of the test DMMU and reference tissue concentrations for the final interpretation “worst case” analyses were based on the adjusted tissue concentrations. The summary PCB tissue data interpretation is provided in **Table 9**.
21. **Bioaccumulation Interpretation**. The DMMP agencies agreed that comparing statistical differences from reference is a necessary, but not sufficient condition to determine a DMMU unsuitable for open-water disposal. For those DMMUs that were statistically greater than reference, a more in depth evaluation was required to determine the significance of the bioaccumulation that had occurred. This evaluation focused on a) Food and Drug Administration (FDA) Action Levels for Poisonous and Deleterious Substances in Fish and Shellfish for Human Food; b) DMMP target tissue concentration value for chemicals of concern to human health.

Table 8. Sediment: Initial and Retested Total PCB Ratio for DMMU-NB-1

| DMMU-NB-1 (10 -14 ft) North Bascule | | | | |
|-------------------------------------|-------|-------------|------------|------------|
| Chemical | Units | Initial (I) | Retest (R) | Ratio: I/R |
| Total PCB (dry weight) | ug/kg | 910 | 180 | 5.1 |
| Total PCBs (TOC normalized) | mg/kg | 91.0 | 24 | 3.8 |
| TOC % | | 1.0 | 0.736 | |

Table 9. Bioaccumulation Testing Results Summary for DMMU-NB-1*

| | | | DMMU_NB-1 (10 - 14 ft) | | | | | | | | | | | |
|---------------|----------|-----------|--------------------------|--------------------------|------------------------|-----------------------|--|-------------------------------|-----------------------|--------------------------|------------------------|-----------------------|--|-------------------------------|
| | | | Macoma nasuta | | | | | | Nephtys caecoides | | | | | |
| | | | DMMU tissue (Initial)(m) | DMMU tissue (unadjusted) | DMMU tissue (adjusted) | Reference (CR-23 Mod) | Statistically different from reference | statistically below guideline | DMMU tissue (Initial) | DMMU tissue (unadjusted) | DMMU tissue (adjusted) | Reference (CR-23 Mod) | Statistically different from reference | statistically below guideline |
| CHEMICAL NAME | Units | Guideline | | | | | | | | | | | | |
| Total PCBs | ug/kg-ww | 750 | 5 (u) | 121 | 617 | 4 (u) | Yes | Yes | 8 (u) | 65 | 332 | 4 (u) | ND | ND |

Note: (1) All tissue concentrations for Total PCBs were analyzed on a wet weight basis to facilitate guideline
 (2) Adjustments to tissue concentrations based on initial sediment versus retested sediment concentration ratios (see Table 8).
 Concentration ratios greater than 1 were adjusted.

ND = Not Determined

Target Tissue Guideline exceeded
NFAR No further Action Required, DMMU Unsuitable for unconfined-open water disposal

*Attachment 1 provides full Bioaccumulation Tissue PCB Aroclor Testing Summary

- a) The FDA guidelines for PCBs is as follows: 2.0 ppm ww
- b) The DMMP TTL for PCBs: 750 ppb (0.75 ppm) ww
22. The DMMP agencies re-evaluated the PCB TTL for human health (December 1999 DMMP Memo attached as **Attachment 5**). Recalculation of the PCB TTL for the Elliott Bay disposal site included using an updated cancer slope factor, recent fish consumption data, and consideration of PCB biomagnification due to trophic transfer. Based on this analysis, an interim TTL for total PCBs (Aroclor) of 750 ppb (0.75 ppm) wet weight was used to interpret bioaccumulation data for the South Bridge Project. These guidelines will likely change in the future as the DMMP review the bioaccumulation testing and TTLs guidance for Puget Sound.
23. The bioaccumulation testing results for both species using sediments from **DMMU-NB-1** were compared to the TTL interpretation guideline. The single DMMU was quantitated less than the TTL for both species, and was subjected to a one-tailed one-sample t-test for the replicated *Macoma* tissue (see **Table 9**). An alpha level (the probability of making a Type I error, rejecting the null hypothesis of no difference between test, reference, and TTL responses when, in fact, they are not different) of 0.1 was selected for these statistical comparisons by the DMMP agencies to reflect the higher within sample variability, and to increase the power of the test to discriminate between reference, TTL and test responses. The test results for *Macoma* were statistically greater than the reference sediment, but below the TTL. Therefore, based on these bioaccumulation testing results **DMMU-NB-1** sediments are suitable for open-water disposal at the Elliott Bay site.
24. **Suitability for Unconfined-Open Water Disposal.** **Table 10** provides the testing results summary for four tested DMMUs (**SB-1, SB-2, NB-1, and NB-2**) amounting to 3,000 cy. DMMU-SB-3 (750 cy) underlying tested DMMU's **SB-1** and **SB-2** was found to be suitable based on best professional judgment, after reviewing the overlying sediment quality results for **SB-1** and **SB-2**. Native material underlying South and North bascules, amounting to 8,000 cy and 11,500cy respectively was found to be suitable based on best professional judgment after reviewing the overlying sediment quality results at both locations. The results summarized in this suitability determination indicate that a total volume of **23,250 cy** is **suitable** for unconfined-open-water disposal at the Elliott Bay non-dispersive site.
25. As summarized in **Table 10**, surface material amounting to a total cumulative volume of **3,057.4 cy** at both bascules (includes 60.4 cy of sand blanket material), which was not tested during this characterization effort, but sediment testing conducted in 2004² were previously reviewed by DMMP, and found to be **unsuitable** for unconfined-open-water disposal (**Attachment 1**). At the South bascule, the top 6 feet of material (cumulative volume = 1,150.2 cy, which also includes removal of 30.2 cy sand blanket), and at the North bascule, the top 10 feet of material (cumulative volume = 1,897.2 cy, which also includes removal of 30.2 cy sand blanket) is unsuitable for open-water disposal, and must be disposed at an Ecology approved upland disposal site.
26. As previously discussed, and summarized in **Table 10**, the 5,415 cy of backfilled upland sourced material will be removed within each bascule cofferdam after bascule pier construction for a total of 10,830 cy. Once the caissons are in place, the backfill material remaining between the caissons and the interior walls of the cofferdams will be removed down to the riverbed prior to removing the cofferdam structures. That volume would represent an additional volume of backfill material of 1,040 cy per pier, or 2,080 cy total.
27. This memorandum documents the suitability testing outcome for the testing conducted for the proposed dredging at the South Park Bridge Project for unconfined-open-water disposal at the Elliott Bay non-dispersive disposal

² Wilbur Consulting. 2004. Final preliminary site investigation report for the South Park Bridge project. Prepared for King County Department of Transportation. Wilbur Consulting, Inc., Seattle, WA

site. However, this suitability determination does not constitute final agency approval of the project. A dredging plan for this project must be completed as part of the final project approval process. A final decision will be made after full consideration of agency input, and after an alternatives analysis is done under Section 404(b)(1) of the Clean Water Act.

Table 10. DMMU Specific and Total Project Testing Outcome Summary (Errata revision: 12/9/10).

| Location | Depth Interval (ft) Below mudline | DMMU ID | Original Estimated Volume (cy) | Revised Volume Estimate (cy) | Suitability Outcome |
|-----------------------------------|--|---------|--------------------------------|---|--|
| South bascule | 0 – 6 | None | 1,100 | 1,150.2 (includes 30.2 cy sand blanket) | Unsuitable based on previous testing results (Wilbur, 2004) |
| | 6 – 10 | SB-1 | 750 | * | Suitable |
| | 10 – 14 | SB-2 | 750 | * | Suitable |
| | 14 – 18 | SB-3 | 750 | * | Suitable (no testing required based on overlying sediment quality results) |
| | 18 – 75 | None | 8,000 | * | Suitable: Native (no testing required) |
| | Cofferdam backfill material excavated for caisson construction | None | Omitted | 5,415 | Unsuitable |
| | Backfill material removed prior to cofferdam removal | None | Omitted | 1,040 | Unsuitable |
| North bascule | 0 – 10 | None | 1,850 | 1,897.2 (includes 30.2 cy sand blanket) | Unsuitable based on previous testing results (Wilbur, 2004) |
| | 10 – 14 | NB-1 | 750 | * | Suitable (based on toxicity and bioaccumulation testing results) |
| | 14 – 18 | NB-2 | 750 | * | Suitable (PCB testing only) |
| | 18 – 100 | None | 11,500 | * | Suitable: Native (no testing required) |
| | Cofferdam backfill material excavated for caisson construction | None | Omitted | 5,415 | Unsuitable |
| | Backfill material removed prior to cofferdam removal | None | Omitted | 1,040 | Unsuitable |
| Total estimated Suitable volume | | | 23,250 cy | * | |
| Total estimated Unsuitable volume | | | 2,950 cy | 3,057.4 cy (includes 60.4 cy sand blanket) | Based on 2004 testing results (Wilbur Consulting, 2004) |
| | | | Omitted | 12,910 cy backfill material | |

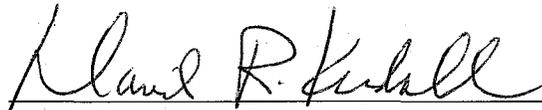
*The estimated volumes are either unchanged or lower than the original volume estimates

SUBJECT: DETERMINATION ON THE SUITABILITY OF SUPPLEMENTAL CHARACTERIZATION OF PROPOSED DREDGED MATERIAL FROM KING COUNTY DEPARTMENT OF TRANSPORTATION SOUTH PARK BRIDGE PROJECT (CORPS APPLICATION: NWS-2009-1586) DREDGING PROJECT IN ELLIOTT BAY, WASHINGTON EVALUATED UNDER SECTION 404 OF THE CLEAN WATER ACT FOR OPEN-WATER DISPOSAL AT A DMMP NON-DISPERSIVE OPEN-WATER DISPOSAL SITE

Concur:

12/20/2010

Date



David R. Kendall, Ph.D., Seattle District Corps of Engineers

12/13/2010

Date



Erika Hoffman, Environmental Protection Agency

12/13/2010

Date



Laura Inouye, Ph.D., Washington Department of Ecology

12/14/2010

Date



Dave Vagt, Washington Department of Natural Resources

Copied furnished:

Jack Kennedy, Corps Regulatory Project Manager

Jim Sussex, King County

Joanna Florer, Windward Environmental

Erika Hoffman, EPA

Laura Inouye, Ph.D. Department of Ecology

Dave Vagt, DNR

DMMO file

Attachment 1. DMMP Coordination on SPB Project

From: Kendall, David R NWS [mailto:David.R.Kendall@usace.army.mil]
Sent: Tuesday, December 16, 2008 2:40 PM
To: Turney, Julia
Cc: WASSON, COURTNEY (DNR); Inouye, Laura (ECY); Hoffman.Erika@epamail.epa.gov
Subject: RE: Dredged Material Management Program Review of the South Park Bridge Foundation Sampling Data
Importance: High

Hi Julia: After reviewing your questions, and coordinating with my DMMP agency counterparts (e.g. I received input back from EPA and Ecology), here is our interpretation on an approach that DMMP would consider for characterization purposes.

1. The 0-10 ft dredging prisms for both North and South Piers would be considered unsuitable with upland disposal, and no DMMP Characterization.

The DMMP Characterization would include the following:

2. Collect a single core sample at each location (e.g., North Pier and South Pier) from 0-18 ft below mudline, with one vertically composited sample collected from the 10-14 ft layer for analysis of DMMP Chemicals of Concern (The full DMMP list + dioxin/furans), and one vertically composited sample from the 14-18 ft layer (archived). Depending on the analysis results of the two 10-14 ft composited samples, the archived samples may be analyzed, if the overlying sample results indicate that sediment quality exceedances of DMMP COC are noted.

3. Collect and archive enough sample from each location (10-14 ft) for toxicity testing if needed.

4. The 18 ft to 75 ft and 18 ft to 100 ft material would be considered "Native" material and not subject to DMMP testing, and would be considered suitable for unconfined-open-water disposal.

5. A Sampling and analysis plan would need to be submitted/approved prior to collecting samples as outlined. I have linked our Chemical of Concern

List and links to our Users Manual, which provide general information regarding sampling requirements.

http://www.nws.usace.army.mil/PublicMenu/documents/DMMO/2007_COCs.pdf

http://www.nws.usace.army.mil/PublicMenu/documents/DMMO/July_2008_UM.pdf

Please let me know if you have further questions.

David

David R. Kendall, Ph.D.

Chief, Dredged Material Management Office Seattle District, Corps of Engineers

Phone: 206/764-3768

Fax: 206/764-6602

email: david.r.kendall@nws02.usace.army.mil

From: Turney, Julia [mailto:Julia.Turney@kingcounty.gov]

Sent: Friday, December 12, 2008 3:22 PM

To: Kendall, David R NWS

Cc: LINO461@ecy.wa.gov; Hoffman, Erika; courtney.wasson@dnr.wa.gov

Subject: RE: Dredged Material Management Program Review of the South Park Bridge Foundation Sampling Data

Hi David,

Thank you for your comments. This feed back is very helpful and we appreciate your offer to work with our consultant to develop the sampling plan. You had asked a question about depth to native sediments and material volumes. I've responded to these questions below and have included a question that was raised during our discussions yesterday. I left you a voice mail about this today.

1) You had asked a question about where "native ground" started. The boring logs for SB-4 and SB-5 in the Duwamish identify the change from fill to alluvial deposits between 10 and 20 feet below ground surface (bgs), 16 feet bgs in Boring SB-5. This depth is consistent with the excavation and dredge records for the Duwamish channel. The sediment density (as determined from blow counts during drilling) increases below 10 feet, indicating that native ground has been reached. You noted the high sediment quality below 50 feet.

This depth is the transition zone between sand-silt sediments and predominately finer silt deposits.

2) The sediment volume estimates for the upper contaminated sediments (historic Duwamish deposits since 1911) would be 70' x 70' x 16' (worst case depth of contamination) for a volume of 2900 cubic yards for upland disposal for each pier. The volume of native sediments below the channel bottom would be 16,151 cubic yards in the north pier. The south pier is expected to be shallower and would generate 9800 cubic yards of native material. The total open water disposal volume would be approximately 26,000 cubic yards of native material for open water disposal.

3) We are planning seven borings, three at one pier location and four at the second pier. Your email noted that at most two samples per boring location would be needed. We wanted to confirm whether you were suggesting numbers of samples per boring, not per pier. The answer to this question will help us with estimating our contract supplement.

Thank you again.

Julia Turney
206-296-0267
julia.turney@kingcounty.gov

From: Kendall, David R NWS [mailto:David.R.Kendall@usace.army.mil]
Sent: Friday, December 05, 2008 2:24 PM
To: Turney, Julia
Cc: LINO461@ecy.wa.gov; Hoffman, Erika; courtney.wasson@dnr.wa.gov
Subject: Dredged Material Management Program Review of the South Park Bridge Foundation Sampling Data
Importance: High

Hi Julia: I wanted to get you my impressions on the data provided relative to DMMP requirements. Unfortunately, because of workload issues, my DMMP agency

counterparts have not had a chance to review and provide input, so the comments I am providing are mine alone. I have linked my colleagues, so they can embellish or provide their individual input to this exchange. The data you provided show a relatively contaminated sediment layer down to 10 feet, especially for PCBs, which will likely not meet our open-water disposal guidelines. The sediment quality data for borings below 10 feet down to 100 feet look much better and generally depict relatively clean sediments.

The detection limit exceedances in SQS noted for Butyl-benzyl-phthalate appear to be largely due to low TOC's in those sediments, ranging from 0.24 to 0.34%. The Butyl-benzyl-phthalate expressed on a dry weight basis are all well below the DMMP Screening Levels (e.g., 63 ppb), and would not be a problem. Characterization of the material below 10 feet, would generally require analysis of all DMMP Chemicals-of-Concern (COC) (http://www.nws.usace.army.mil/PublicMenu/documents/DMMO/2007_COCs.pdf). The results reported for the samples below 10 feet do not include PCBs, and ionizable organic compounds (phenol, 1-methyphenol, 2,4-dimethylphenol, pentachlorophenol, Benzyl alcohol, Benzoic acid), which would need to be quantified for a DMMP evaluation.

The relatively high sediment quality in the deeper sampling depths (>50 feet) lead me to ask the question, do the geotechnical boring logs note the "Native" sediment contact layer? We can consider giving Native sediments a pass without sediment testing requirements, but generally would require analysis of the first 4-8 feet of that layer to demonstrate the sediment quality as part of the DMMP evaluation.. You indicated in your earlier email, that the total volume to be removed and managed is 31,758 cy.

It would be helpful to know what the volume is of the top 7-10 feet of material, that appears to be heavily contaminated with PCBs, and the volume of the remaining subsurface material including Native sediments? Given the volume stipulated, which is less than 31,000 cy with the surface sediment removal, the volume to be characterized within each subsurface location is probably less than 15,000 cy, which under a Moderate ranking, would be a, minimum of one analysis at each location. Therefore, the likely testing that would be required at both locations to evaluate the subsurface sediment quality should be at most no more than 2 analyses per boring location. We would be happy to work with you or your contractor to scope out a sampling/testing strategy that meets our needs. Please

call me if you have any questions about my analysis of likely DMMP testing requirements for your project.

David

David R. Kendall, Ph.D.

Chief, Dredged Material Management Office Seattle District, Corps of Engineers

Phone: 206/764-3768

Fax: 206/764-6602

email: david.r.kendall@nws02.usace.army.mil

**SOIL BORINGS SB4 and SB5
DATA SUMMARY REPORT**

**South Park Bridge Replacement Project
Seattle, Washington**

Prepared by

King County Department of Transportation
Road Services Division
201 S Jackson Street
Seattle, WA 98104

November 12, 2008

ACRONYMS

SUMMARY

FIGURES

- 1 Project Vicinity Map
- 2 Boring Location Map

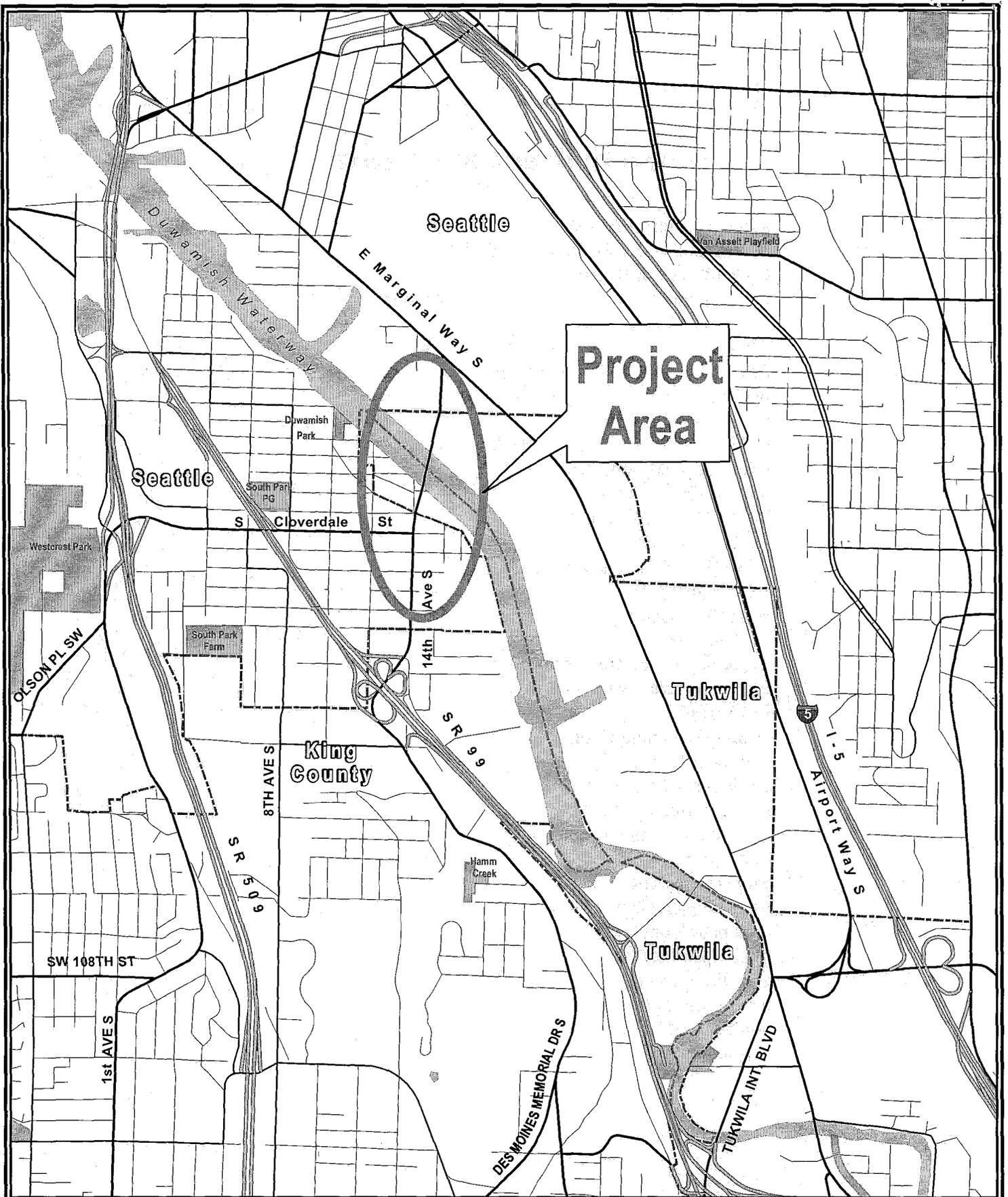
TABLES

Summary of Chemical Results for SB-4 and SB-5 Samples Collected for the South Park Bridge Project, June through August 2003.

REFERENCES

Abbreviations and Acronyms

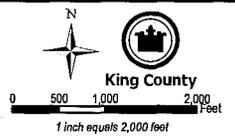
| | |
|----------|---|
| ACM | asbestos-containing materials |
| ASTM | American Society for Testing and Materials |
| CFR | Code of Federal Register |
| cPAHs | carcinogenic polycyclic aromatic hydrocarbons |
| DSOA | Duwamish Sediment Other Area |
| Ecology | Washington State Department of Ecology |
| EIS | Environmental Impact Statement |
| LBP | lead-based paint |
| LDW | Lower Duwamish Waterway |
| LUST | leaking underground storage tank |
| mg/kg | milligrams per kilogram |
| µg/L | micrograms per liter |
| MTCA | Model Toxics Control Act |
| MW | monitoring well |
| PCB | polychlorinated biphenyl |
| PCE | perchloroethylene |
| PP | priority pollutant |
| ppm | parts per million |
| PSDDA | Puget Sound Dredge Disposal Analysis |
| RCRA | Resource Conservation and Recovery Act |
| SB | soil boring |
| CSL | cleanup screening level |
| SMS | Sediment Management Standard |
| SQG | small quantity generator |
| SQS | sediment quality standard |
| SVOC | semi-volatile organic compound |
| TBT | tri-butyl tin |
| TCE | trichloroethylene |
| TEF | toxicity equivalent factor |
| TPH | total petroleum hydrocarbons |
| TOC | total organic carbon |
| U.S. EPA | U.S. Environmental Protection Agency |
| UST | underground storage tank |
| VOC | volatile organic compound |
| WAC | Washington Administrative Code |



Project Vicinity Map

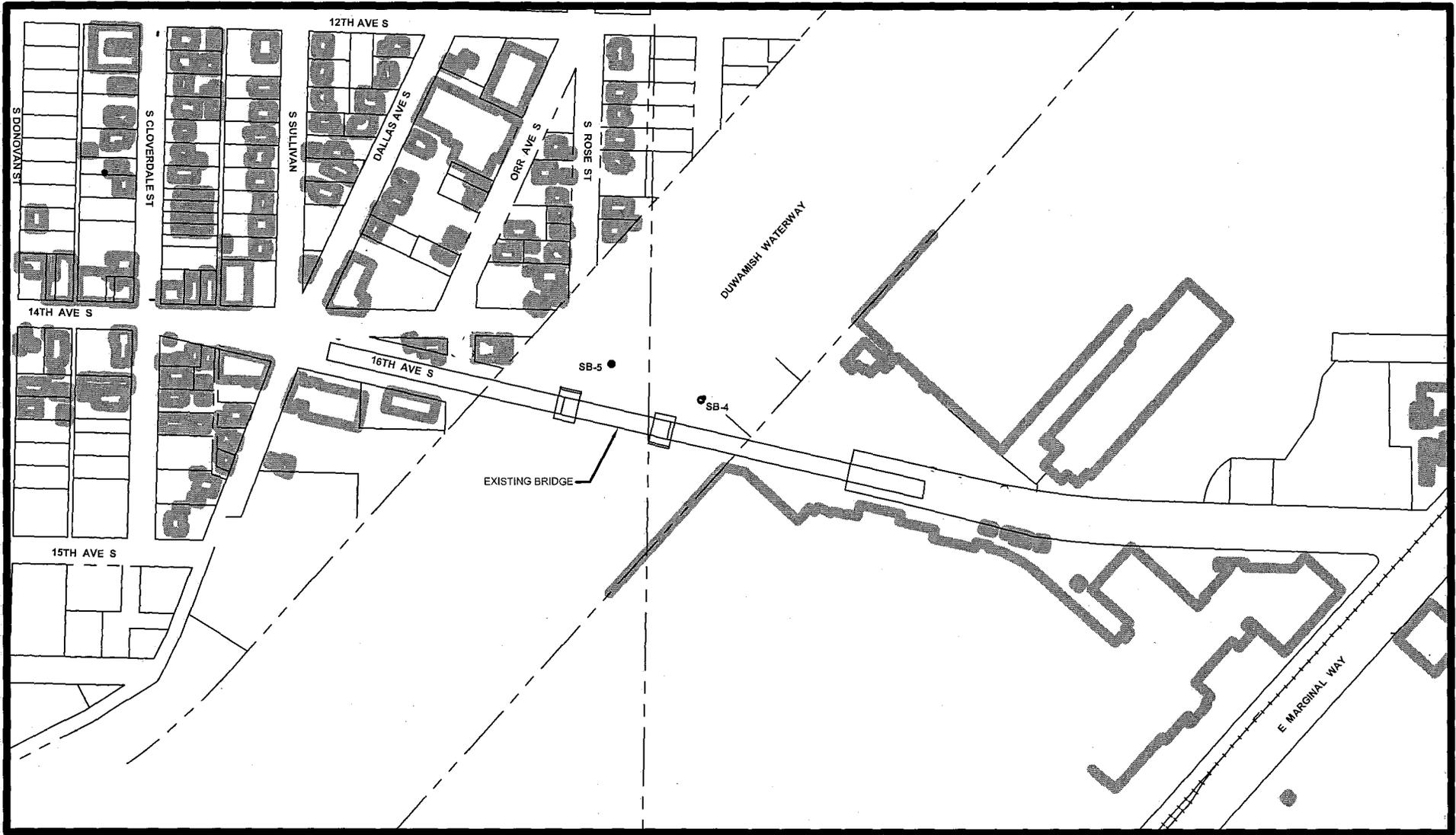
-  Hydro
-  Parks
-  City Boundary

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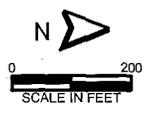


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Figure 1



DATE: 09/25/04 3:32pm FILE: BL1921063P05T01-402



- LEGEND**
- Soils Borings
 - Monitoring Well

Figure 2
Boring Location Map

2003 Sediment Borings SB-4 and SB-5 (Herrera, 2008)

Borings SB-4 and SB-5 were drilled approximately 5 meters (16.4 feet) downstream of the existing bridge, with SB-4 near the north riverbank and SB-5 near the south riverbank of the waterway (Figure 2). Sediment samples were collected from both borings, extending down from the water/mud line interface to maximum depths of 100 feet (SB-4) and 75 feet (SB-5). A total of 11 sediment samples were analyzed for Puget Sound Dredge Disposal Analysis (PSDDA) volatiles, PSDDA SVOCs, PSDDA pesticides/PCBs, metals, total organic carbon (TOC), and tri-butyl tin (TBT). Complete laboratory analytical reports, including sample chain-of-custody forms, and laboratory quality assurance/quality control documentations of all sediment samples collected from SB-4 and SB-5 are presented on a CD-ROM in Appendix C of the 2003 preliminary site investigation report (Wilbur 2004b).

Analytical results of these sediment samples are summarized in the attached tables and include the following:

- Diesel- and lube oil-range hydrocarbons were detected in sediment samples collected from both borings at depths between 2.5 and 10 feet below the mud line. No SQS or CSL are established for diesel- and lube oil-range hydrocarbons in sediment. Diesel concentrations ranged between 19 and 130 mg/kg (dry weight) and lube oil concentrations ranged between 22 and 230 mg/kg (dry weight).
- Low concentrations of five VOCs, including methylene chloride, acetone, carbon disulfide, 2-butanone, and 4-isopropyltoluene, were detected in all 11 sediment samples collected from SB-4 and SB-5. No SQS or CSL are established for these VOC constituents in sediment.
- A mixture of PCBs (Aroclors 1248, 1254, and 1260) were identified in six of the 11 samples collected from both borings extending down from the mud line to 10 feet below the mud line; total PCB concentrations exceeded either the SQS criterion (12 mg/kg organic carbon [OC]) or the CSL criterion (65 mg/kg OC) in all six samples.
- Butyl benzyl phthalate was detected in the sediment sample collected at the 7-foot interval below the mud line from SB-4, with the concentration exceeding the SQS criterion of 4.9 mg/kg OC, but below the CSL criterion of 64 mg/kg OC. Bis(2-ethylhexyl)phthalate was detected in samples collected at 5 feet below the mud line from SB-4 and 75 feet below the mud line from SB-5; both concentrations exceeded the CSL criterion (78 mg/kg OC). Other SVOC constituents were either detected below SMS criteria, or not detected.

- Between four and ten metals were detected in all 11 sediment samples from SB-4 and SB-5. None of the metals were detected above applicable SMS criteria. No TBT was detected in any of the samples collected.

Sediment Analytical Results Table –
2003 Preliminary Site Investigation and
Lower Duwamish Waterway Group
2003 Sediment Quality Database

Attachment 2. Mean percent survival of *Macoma nasuta* & *Nephtys caecoides* during 45-day bioaccumulation test.

| Treatment | <i>Macoma nasuta</i> | | <i>Nephtys caecoides</i> | |
|------------|----------------------|---|--------------------------|---|
| | Mean \pm S.D. | n | Mean \pm S.D. | n |
| Control | 88.3 \pm 4.6 | 5 | 99.0 \pm 2.2 | 5 |
| Reference | 81.7 \pm 10.9 | 5 | 98.0 \pm 2.7 | 5 |
| NB-1014-BT | 90.0 \pm 7.0 | 5 | 99.0 \pm 4.2 | 5 |

SD = Standard Deviation

Attachment 3. Initial and final mean biomass of *Macoma nasuta* & *Nephtys caecoides* used in 45-day bioaccumulation test.

| Treatment | <i>Macoma nasuta</i> | | <i>Nephtys caecoides</i> | |
|------------|----------------------|----|--------------------------|----|
| | Mean \pm S.D. | n | Mean \pm S.D. | n |
| Day 0 | 4.41 \pm 1.70 | 15 | 0.79 \pm 0.27 | 15 |
| Day 45: | | | | |
| Control | 3.82 \pm 0.87 | 15 | 0.40 \pm 0.16 | 15 |
| Reference | 3.28 \pm 0.84 | 15 | 0.53 \pm 0.25 | 15 |
| NB-1014-BT | 3.31 \pm 0.79 | 15 | 0.49 \pm 0.17 | 15 |

Proposed Approach for Developing an Interim Target Tissue Level for Total PCBs Based on Risk to Human Health

Prepared by DMMP Agencies – Seattle, WA
December 21, 1999

Introduction

The Dredged Material Management Program has developed a site-specific, interim target tissue level (TTL) for total PCBs (tPCBs)¹ in benthic organisms based on a recalculation of the TTL used in the March 1997 suitability determination for the Port of Seattle Terminal 18 (95-02133). The human health risk assessment used to derive this interim TTL considers consumption of bottom fish only and uses parameters specific to the Elliott Bay dredged material disposal site. Newly available seafood ingestion rate data for high-end consumers (Native American tribes and Asian/Pacific Islanders) and the estimates of the biomagnification potential of tPCBs between bottom fish and their benthic prey were also used. The interim TTL will be used to determine the suitability of Dredged Material Management Units from the East Waterway Stage II project for disposal at the Elliott Bay site based on statistical comparison to tissue data from laboratory bioaccumulation testing. The TTL is considered interim pending incorporation of new seafood consumption rate information and/or the DMMP adopting different approaches to tPCB measurement and toxicity summation.

This memorandum discusses:

1. Information used to estimate cancer risks associated with human exposure to tPCBs derived from the disposal site,
2. The basis and protectiveness of the interim TTL, and
3. Results of applying the interim TTL to the East Waterway bioaccumulation data.

Human Health Risk Assessment

Concentrations of tPCBs in whole body bottom fish (e.g., English sole) associated with a 10^{-5} risk of excess cancer were estimated for different consumers using the following equation:

$$\text{BF tPCBs} = \frac{(\text{Risk}) (\text{BW}) (\text{AT})}{(\text{BF IR}) (\text{BF fr PS}) (\text{BF HR/ DS}) (\text{EF}) (\text{ED}) (\text{SF})(\text{CF1})(\text{CF2})}$$

where:

BF tPCBs = Estimated concentration of tPCBs in a bottom fish (ug/kg wet weight whole body)

Risk = Excess cancer risk of 1×10^{-5} (unitless)

BW = Body weight (70 kg)

AT = Averaging time (70 years (25,550 days))

BF IR = Ingestion rate of bottom fish (g/day)

BF HR = Home range of a bottom fish (2,334 acres from PSDDA, 1988)

BF fr. PS = Fraction of bottom fish consumed that are from Puget Sound (unitless)

DS = Area of the Elliott Bay disposal site (395 acres)

EF = Exposure frequency (365 days/year)

ED = Exposure duration of consumer group (years)

SF = Cancer slope factor for tPCBs (2.0 mg/kg-day)

CF1 = Conversion factor (kg/g)

CF2 = Conversion factor (mg/ug)

¹ Total PCBs are currently calculated as the sum of the *detected* Aroclor concentrations.

The following assumptions were used in estimating bottom fish tissue concentrations:

- Bottom fish (e.g., English sole) are the only type of seafood consumed that could be exposed to sediment-associated tPCBs at the Elliott Bay disposal site.²
- Bottom fish reach the calculated body burden of tPCBs only from exposure to contaminated benthic prey from the Elliott Bay disposal site.

Exposure parameters and bottom fish tPCBs concentrations are presented in Tables 1 and 2. Fish tissue concentrations were estimated using ingestion rate information for tribal (Toy *et al.*, 1996), Asian/Pacific Islander (EPA, 1999), and recreational consumers (Landolt *et al.*, 1985). The following parameters varied according to consumer group:

- Percentile ingestion rate of bottom fish among all fish consumed
- Fraction of bottom fish eaten that are caught in Puget Sound (versus eaten at a restaurant or bought at a market)
- Exposure duration (e.g., 30 years for recreational fishers and Asian/Pacific Islanders, 70 years for tribal consumers).³

An overview of the excess cancer risks for different consumer populations associated with various concentrations of tPCBs in fish is presented in Table 3. The range of tPCB tissue concentrations evaluated (341 ppb - 7531 ppb wet weight) correspond to the 10^{-5} risk levels for the different consumer populations as calculated in Tables 1 and 2.

Biomagnification

Biomagnification of tPCBs has been widely observed to occur between different trophic levels with increasing concentrations observed for higher level consumers. Particularly high biomagnification factors (BMF) have been observed between benthic/epibenthic organisms and bottom feeding fish (e.g., Metcalfe and Metcalfe, 1997). The most recent TTL for tPCBs (2.0 ppm wet weight) calculated for the Port of Seattle T-18 project did not take biomagnification into account. However, in recalculating this interim TTL we have used a BMF to relate the acceptable concentration in a bottom fish to that in the benthic invertebrates exposed to the sediments.

The agencies have reviewed the available data on concentrations of tPCBs in the tissues of bottom fish and their prey in the Harbor Island/Elliott Bay area and have concluded that this is insufficient for use in calculating a site-specific BMF. For the purposes of developing an interim TTL for the East Waterway Stage II project, the agencies have decided to use a non-site-specific BMF. In consultation with Phil Cook (EPA/ORD, Duluth), a factor of 2 was chosen as a reasonable estimate of the BMF for tPCBs between benthic organisms and bottom feeding fish (whole body basis). This estimate falls within the range of 2 - 4 reported by Metcalfe and Metcalfe (1997) for benthic feeders (sucker and sculpin) and is similar to the value of 2.7 used by the New York/New Jersey Harbor Dredging Forum for summer flounder based on a two-step trophic model originally elaborated by Frank Gobas (Zambrano, 1993).

² PSDDA deep-water disposal sites such as the one in Elliott Bay were originally selected to avoid fishery areas, particularly areas where high concentrations of shellfish may be found.

³ The DMMP agencies have decided to use a generic exposure time of 70 years for tribal consumers based on assumed patterns of tribal residence. Native Americans, wishing to maintain cultural ties, may relocate over a limited geographic area and continue to visit their usual and accustomed fishing areas. Furthermore, subsistence anglers may share their catch with their families, increasing the effective exposure duration for family members. Efforts to obtain regional data on relocation or duration of residence for local tribes have not been successful. The agencies will reconsider use of the 70 year exposure duration if and when data is available indicating that this is value is overly conservative.

Calculation of TTLs

Earlier PSDDA assessments as well as remediation projects for tPCBs in Puget Sound (e.g., Manchester Superfund) have used an upper limit of 1×10^{-5} excess cancer risk in deriving acceptable concentrations of tPCBs in fish tissue. Based on information provided in Table 3, tPCB concentrations of approximately 600 ppb wet weight in bottom fish exposed to the Elliott Bay disposal site would be protective at the 1×10^{-5} risk level for the 90th percentile consumption rate of tribal consumers. Tissue tPCB concentrations of approximately 1500 ppb wet weight would be similarly protective for the general population of Asian/Pacific Islanders. PCB concentrations in fish that are protective at the 1×10^{-5} risk level for recreational fishers are considerably higher, falling between 4000 - 7500 ppb depending on the ingestion rate used (Table 2).

As discussed in the preceding section, an estimated biomagnification factor of 2 was used to convert the tPCB concentration in a bottom fish to that in a trophically-linked benthic organism. Using the acceptable fish tissue concentrations for both tribal and Asian/Pacific Islander groups results in TTLs of 300 and 750 ppb wet weight, respectively. Using acceptable fish tissue concentrations for recreational consumer results in TTLs ranging from 2000 to 3750 ppb wet weight (depending on the ingestion rate used).

Selection of an Interim TTL for tPCBs

After considering the information presented in Table 3, the DMMP agencies concluded that an interim TTL for tPCBs based on risk to recreational consumers would not be suitably protective of high end consumers represented by tribes and Asian/Pacific Islanders. However, use of the most protective TTL value (300 ppb wet weight) based on 1×10^{-5} risk to tribal consumers may be overprotective, as conservative exposure parameter values were assumed in order to compensate for uncertainties that might underestimate risk. Therefore, the agencies have qualitatively evaluated the extent to which the following assumptions over- and under- estimated exposure (the DMMP's view of the influence on the risk estimate is indicated in parenthesis):

Over-protective assumptions:

- Calculations of risk are based on high-end (tribal and Asian/Pacific Islander) consumption rates rather than those of recreational fishers. (Important influence)
- The fraction of seafood harvested from all of Puget Sound is used in this calculation to represent the fraction of seafood harvested from an area influenced by the Elliott Bay disposal site. This value likely overestimates the fraction harvested from the Elliott Bay. (Important influence)
- Food preparation practices or cooking methods that might reduce tPCB concentrations were not considered in the evaluation. (Important influence)
- The Elliott Bay disposal site is assumed, for the sake of this evaluation, to be uniformly covered with the PCBs from each separately evaluated dredged material management unit. However, each management unit would be in fact mingled with others during physical placement of dredged material at the site, resulting in a *site* concentration, which is lower than many of the management units of concern. Prey items for bottom fish are thus assumed to have uniformly higher tissue concentrations of tPCBs than would be expected. (Possibly important influence)
- Assumption of a 70 year exposure period for tribal consumers is based on patterns of tribal residence rather than site-specific information. (Moderate influence)
- We did not use a whole body/fillet factor to account for the difference in lipid content between the whole fish (higher lipid) and what is typically considered to be the edible portion of a fish (lower lipid) in calculating the interim TTL. Although no specific information is available, it appears culinary practice of some Asian/Pacific Islanders may involve eating the

entire fish rather than fillets. Hydrophobic organic compounds (such as tPCBs) tend to preferentially concentrate in the lipids of aquatic organisms. (Unknown influence)

Under-protective assumptions:

- The 90th percentile Asian/Pacific Islander consumption rates used included non-fish consumers and hence would underestimate the actual bottom fish consumption rate. (Important influence)
- There are scant data available to calculate 90th percentile bottom fish consumption rates for individual ethnic populations that comprise the general category of Asian/Pacific Islanders. Thus the consumption rate for API may not be protective of all individual populations. (Possibly important influence)
- Estimated tissue concentrations of tPCBs in bottom fish are assumed to come solely from their exposure to the disposal site (i.e., existing tPCB body burdens in bottom fish from sources other than the disposal site are not considered). (Minor influence⁴)
- Human consumer exposure to Elliott Bay disposal site tPCBs occurs solely through ingestion of bottom fish (i.e., assumes no exposure from eating shellfish or pelagic fish that could pick up tPCBs from disposal site). (Likely to be a minor influence - see footnote #2)
- Non-cancer adverse health effects of tPCBs to human consumers are not considered in risk calculations; cancer risks only are considered. (Unknown but probably minor influence)

Based on the foregoing considerations, the DMMP agencies have concluded that that the assumptions about exposure used in this assessment tend to overestimate the actual exposure. **Thus, the agencies have selected 750 ppb wet weight as the interim TTL for tPCBs in benthic organisms.** The excess cancer risk associated with this interim TTL is 9.7×10^{-6} for the general population of Asian/Pacific Islander consumers and 2.6×10^{-5} for tribal consumers. The DMMP agencies consider the calculated upper risk limit associated with this value to be acceptable for the purposes of deriving an interim TTL for the Elliott Bay dredged material disposal site.

Comparison of Proposed TTL to results of Bioaccumulation Testing

Bioaccumulation data from the East Waterway testing are presented in Appendix 8. Tissue concentrations are corrected for differences between round 1 and round 2 sediment tPCB concentrations. The results of statistical comparisons between the corrected tissue concentrations and the interim TTL of 750 ppb wet weight are indicated. Application of the tPCB TTL results in failure of 3 DMMUs (S-11, S-16, S-23) out of a total of 13 DMMUs tested for tPCBs. A total of 25 DMMUs were tested for bioaccumulation out of 99 DMMUs that were evaluated during the Stage II testing (no bioaccumulation testing was performed on the 8 DMMUs tested from USCG Slip 36 dredging area).

⁴ Based on comparison to fish tissue data for Elliott Bay indicating that average tissue concentrations of total PCBs in English sole range from 40-70 ppb wet weight (EVS Solutions, 1999).

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Table 1. Calculation of Allowable Bottom Fish Tissue Concentrations Based on Risk to Tribal and Asian/Pacific Islander Consumers

| Risk ¹ | Body wt. (kg) | Avg. time ² (days) | BF ³ ingested (g/day) | | BF fraction from PS ⁴ | | Disposal area / BF home range ⁵ | Exp. ⁶ freq (days/yr) | Exp. time (years) | Slope Factor mg/kg-day | Calc. Fish PCB conc. ug/kg wet wt. | Source of ingestion rate and BF fraction used |
|-------------------|---------------|-------------------------------|----------------------------------|---|----------------------------------|---|--|----------------------------------|-------------------|------------------------|------------------------------------|--|
| 0.00001 | 70 | 25,550 | 18.3 | a | 0.17 | b | 0.169 | 365 | 30 | 2 | 1,553 | All 90th %ile Asians/Pacific Islanders (EPA, 1999) |
| 0.00001 | 70 | 25,550 | 18.5 | c | 0.13 | d | 0.169 | 365 | 70 | 2 | 861 | Squaxin 90th%ile Tribal (from Toy et al., 1996) |
| 0.00001 | 70 | 25,550 | 14.3 | e | 0.39 | f | 0.169 | 365 | 70 | 2 | 371 | Tulalip |
| 0.00001 | 70 | 25,550 | 17.4 | g | 0.21 | h | 0.169 | 365 | 70 | 2 | 567 | weighted mean |

¹ Corresponds to 1 additional cancer per 100,000 population

² Equivalent to a life expectancy of 70 years.

³ Bottom Fish

⁴ Puget Sound

⁵ 395 acres / 2335 acres = 0.169

⁶ Exposure

a. 90th %ile consumption rate of bottom fish by Puget Sound Asians and Pacific Islanders (n=202) (EPA, 1999).

Individuals that do not consume bottom fish may be included in consumption rate calculation (Kissinger, personal communication, 1999).

b. Weighted mean of 17% of the bottom fish ingested by Asians and Pacific Islanders interviewed were caught in Puget Sound Waters (vs. purchased) (EPA, 1999).

c. 90th %ile consumption rate of bottom fish by the Squaxin tribe (n=85). Values from Toy et al (1996) adjusted to exclude non-bottom fish consumers (Kissinger, personal communication, 1999).

d. A mean of 13% of the bottom fish ingested by Squaxin tribal members interviewed were caught in Puget Sound Waters (Toy et al., 1996).

e. 90th %ile consumption rate of bottom fish by Tulalip tribal members (n=34). Values from Toy et al (1996) adjusted to exclude non-bottom fish consumers (Kissinger, personal communication, 1999).

f. A mean of 39% of the bottom fish ingested by Tulalip tribal members interviewed were caught in Puget Sound Waters (Toy et al., 1996).

g. 90th %ile weighted consumption rate of bottom fish by both the Squaxin and Tulalip tribes (n=119).

Values from Toy et. al (1996) adjusted to exclude non-bottom fish consumers (Kissinger, personal communication, 1999).

h. Weighted mean of 21% of the bottom fish ingested by Squaxin and Tulalip tribal members interviewed were caught in Puget Sound Waters (Toy et al., 1996).

Table 2. Calculation of Allowable Bottom Fish Tissue Concentration Based on Risk to Recreational Fishers

| Risk ¹ | Body wt. (kg) | Avg. time ² (days) | BF ³ ingested (g/day) | BF fraction from PS ⁴ | Disposal area / BF home range ⁵ | Exp. ⁶ freq (days/yr) | Exp. time (years) | Slope Factor mg/kg-day | Calc. Fish PCB conc. ug/kg wet wt. |
|-------------------|---------------|-------------------------------|----------------------------------|----------------------------------|--|----------------------------------|-------------------|------------------------|------------------------------------|
| 0.00001 | 70 | 25,550 | 31 a | 0.039 b | 0.169 | 365 | 30 | 2 | 3997 |
| 0.00001 | 70 | 1 | 11 c | 0.025 d | 0.169 | 1 | 1 | 7.7 c | 1956 f |
| 0.00001 | 70 | 1 | 11 c | 0.025 d | 0.169 | 1 | 1 | 2 g | 7531 h |

¹ Corresponds to 1 additional cancer per 100,000 population.

² Equivalent to a life expectancy of 70 years.

³ Bottom Fish

⁴ Puget Sound

⁵ 395 acres / 2335 acres = 0.169

⁶ Exposure

a. Median fish consumption rate by recreational fishers based on Landolt et al.1985.

b. Bottom fish represent 3.9% of the seafood caught by recreational anglers.

Mean of data from Landolt et al. (1985) and Simmonds et al. (1998) as reported in EVS (1999).

c. Average daily seafood consumption rate for seafood caught in urban bays by recreational anglers from Landolt et al. (1985).

d. Bottom flatfish represent 2.5% (by weight) of the seafood caught by recreational anglers according to Landolt et al. (1985).

e. Old cancer slope factor for PCBs.

f. The TTL calculation performed for the Port of Seattle T-18 suitability determination (1997, 95-02133) .

did not consider averaging time, exposure frequency, or exposure duration and used the old slope factor (7.7) for PCBs.

g. Updated cancer slope factor for PCBs.

h. Recalculation of the T-18 TTL using the updated slope factor for PCBs.

Table 3. Estimated Risk From Ingestion of PCB-Contaminated Bottom Fish Associated with the Elliot Bay Disposal Site

| <i>Allowable tPCB in bottom fish ug/kg wet wt.</i> | <i>Bottom fish ingestion rates (g/day)</i> | | | |
|--|---|--|--------------------------|------------------------|
| | Recreational Mean IR from T-18 SD ¹ 0.28 | Mean IR from WWY RA ² 1.21 | Tribal 90% IR 17.4 | A/PI 90% IR 18.3 |
| 7531 | 1.0E-05 | 1.9E-05 | 1.3E-04 | 4.8E-05 |
| 3997 | 5.3E-06 | 1.0E-05 | 7.1E-05 | 2.6E-05 |
| 1553 | 2.1E-06 | 3.9E-06 | 2.7E-05 | 1.0E-05 |
| 1500 ⁵ | 2.0E-06 | 3.8E-06 | 2.6E-05 | 9.7E-06 |
| 567 | 7.5E-07 | 1.4E-06 | 1.0E-05 | 3.6E-06 |
| 341 | 4.5E-07 | 8.5E-07 | 6.0E-06 | 2.2E-06 |

¹ Bottom fish ingestion rate used in March 1997 suitability determination for the Port of Seattle Terminal 18 (95-02133) calculated using an average daily seafood consumption rate (11 g day) and assuming that 2.5% of the fish caught by recreational anglers are bottom fish (both from Landolt et al., 1985). $11 \times 0.025 = 0.28$

² Bottom fish ingestion rate used in West Waterway Human Health Risk Assessment (EVS, 1999) calculated using a median fish consumption rate by recreational fishers (31 g/day) and assuming that bottom fish represent 3.9% of the seafood caught by recreational anglers (Landolt et al., 1985 and Simmonds et al. 1998). $31 \times 0.039 = 1.21$

³ 90th percentile of weighted mean tribal bottom fish consumption rate from Toy et al. (1996).

⁴ 90th percentile consumption rate by Asians and Pacific Islanders from EPA (1999).

⁵ Bolded fish tissue concentration and associated risk estimates were used in calculating interim TTL for East Waterway Project

$$\text{Risk} = [(\text{SF}) \times (\text{fish tissue PCB}) \times (\text{bottom fish IR}) \times (\text{bottom fish home range/site size}) \times (\text{EF}) \times (\text{ED}) \times (0.001\text{kg/g}) \times (0.001 \text{ mg/ug}) / (\text{body wt.}) (\text{AT})]$$

Where: SF = Slope factor; IR = ingestion rate; exp. = exposure; freq. = frequency; EF = exposure frequency; ED = Exposure duration